### RESEARCH ARTICLE



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## Population genetic structure of the globally introduced big-headed ant in Taiwan

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### **Abstract**

Global commerce and transportation facilitate the spread of invasive species. The African big-headed ant, Pheidole megacephala (Fabricius), has achieved worldwide distribution through globalization. Since the late 19th century, Taiwan has served as a major seaport because of its strategic location. The population genetic structure of P. megacephala in Taiwan is likely to be shaped by international trade and migration between neighboring islands. In this study, we investigated the population genetics of P. megacephala colonies sampled from four geographical regions in Taiwan and elucidated the population genetic structures of P. megacephala sampled from Taiwan, Okinawa, and Hawaii. We observed a low genetic diversity of P. megacephala across regions in Taiwan. Moreover, we noted low regional genetic differentiation and did not observe isolation by distance, implying that long-distance jump dispersal might have played a crucial role in the spread of P. megacephala. We sequenced the partial cytochrome oxidase I gene and observed three mitochondrial haplotypes (TW1-TW3). TW1 and TW3 most likely originated from populations within the species' known invasive range, suggesting that secondary introduction is the predominant mode of introduction for this invasive ant. TW2 represents a novel haplotype that was previously unreported in other regions. P. megacephala populations from Taiwan, Okinawa, and Hawaii exhibited remarkable genetic similarity, which may reflect their relative geographic proximity and the historical connectedness of the Asia-Pacific region.

#### KEYWORDS

bridgehead effect, genetic bottleneck, invasion biology, invasive ant, population structure

#### TAXONOMY CLASSIFICATION

Entomology

## INTRODUCTION

Numerous case studies have reported that newly introduced ant species become established invaders outside their native ranges, even if they experience a substantial genetic bottleneck during colonization, which may hinder their fitness (Suarez & Tsutsui, 2008; Tsutsui et al., 2000). This characteristic suggests that the consequences of reduced genetic diversity are not always negative. The apparent

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paradox is that the absence of intraspecific aggression between nests leads to unicoloniality, which contributes to ecological dominance within its introduced range (Drescher et al., 2010; Hoffmann, 2014; Tsutsui et al., 2000). Furthermore, inbreeding in unicolonial invasive ants may purge deleterious alleles, enabling colonies to have an improved chance of survival in new environments (Eyer et al., 2018).

Globalization has facilitated human-mediated biological invasion (Hulme, 2009; Meyerson & Mooney, 2007). With the expansion of international trade and advancements in transportation, the introduction of invasive species to new locations has become increasingly common. This phenomenon is termed the "bridgehead effect" or "secondary introduction" and occurs when one invasion engenders another invasion (Bertelsmeier et al., 2018; Ficetola et al., 2008; Garnas et al., 2016; Lombaert et al., 2010). This process is facilitated by the rapid evolution of invasion-associated traits, including highly plastic life history traits and reduced inbreeding pressure in the intermediate regions, which may increase the propagule pressure of an invader in their introduced regions (Lee, Weng, et al., 2020). The development of high-resolution genetic and genomic markers has facilitated comprehensive analyses of invasion history; such analyses have demonstrated that secondary introduction is the primary mode of introduction for numerous global invasive species, including ants (Ascunce et al., 2011; Blumenfeld et al., 2021; Sherpa et al., 2017).

The African big-headed ant (*Pheidole megacephala*) is an invasive ant species native to Africa and was introduced to most of the world's temperate and tropical zones (Wetterer, 2012). Similar to most invasive ant species (Hee et al., 2000; Tay et al., 2014), *P. megacephala* may not require a large propagule size for successful establishment because a founding propagule comprising 1 queen and at least 10 workers or pupae is sufficient to ensure colony survival (Chang, 1985). *P. megacephala* are habitat generalists, preferring to nest in soil and dead tree logs, which enables them to be transported readily with exported commodities and logs or timber (Sarnat et al., 2015). *P. megacephala* colonies are ecologically and competitively dominant in most areas to which they are introduced, displacing native vertebrates and invertebrates (Burwell et al., 2012; Callan & Majer, 2009; Dejean et al., 2007, 2008; Hoffmann et al., 1999; Plentovich et al., 2009; Strohecker, 2012; Vanderwoude et al., 2000).

In Taiwan, the dominance of *P. megacephala* found in the forest edge has substantially contributed to the collapse of the ant community and species interaction network in the forest (Tsai, 2019). The competitive exclusion of other ants from the forest may decrease the forest interior by at least 1 km from its edge (Tsai, 2019).

In the late 19th century, *P. megacephala* was documented in Africa, the Indian Ocean islands, the Atlantic islands, East Asia, Australia, Hawaii, South America, Central America, and the West Indies (Wetterer, 2012). By the 20th century, this ant species had spread to numerous nearby islands in the Pacific region, such as Hawaii and Australia (Wetterer, 2007). Although their presence in Asia-Pacific countries was documented, these ants' phylogenetics and population genetics received little attention.

To fill the aforementioned research gap, the present study used mitochondrial DNA (mtDNA) and microsatellite DNA to examine the population genetic structure of P. megacephala in Taiwan in order to understand demographic events (e.g., genetic bottleneck and rapid expansion) after their invasion. We inferred the ant population dynamics and postinvasion dispersal pattern in Taiwan. Moreover, we determined whether P. megacephala populations in Taiwan share a common origin with one or more native/introduced populations that share the same patterns as populations in the United States and Australia. Conversely, P. megacephala populations may have arisen multiple times from an unidentified origin, considering that research on the phylogenetics of the test species is limited. Accordingly, we investigated the genetic relationships among the P. megacephala populations from Taiwan, those from two Pacific islands, and those with genetic data on GenBank (including native and several selected introduced ranges).

### 2 | MATERIALS AND METHODS

## 2.1 | Sample collection

A total of 30 *P. megacephala* colonies were sampled from urban parks in four Taiwanese regions (six parks in Taipei [TP], eight in Taichung [TC], eight in Kaohsiung [KH], and eight in Hualien and

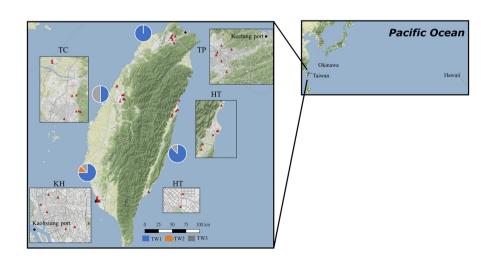


FIGURE 1 Locations of 30 colonies sampled from four administrative regions in Taiwan (red triangle). Black circles indicate Kaohsiung and Keelung, the two major seaports constructed in 1858 and 1913, respectively. They have remained Taiwan's international commerce centers. All three haplotypes with different frequencies were recovered in the sampling regions. Populations sampled from Okinawa and Hawaii were also included for analysis because they are probably historically linked to the populations in Taiwan.

Taitung [HT]; Figure 1, Appendix S1) during 2018–2019. Workers collected from each colony were transported to the laboratory for aggression tests to ensure that they originated from different colonies. Our preliminary study revealed that colonies located more than 100m apart acted aggressively. Furthermore, 6 and 13 colonies of this species were collected from Okinawa and Hawaii, respectively (Appendix S1). All samples used in this study were minor workers. The workers were preserved in 95% alcohol and refrigerated at 4°C until DNA extraction. Morphological identification was based on the methods of Bolton (1994), Lin (1998), and Sarnat et al. (2015). Our preliminary result indicated no cryptic species in our sampling areas (Liu, 2020; Appendix S2). This finding supports that of Wills et al. (2014), who reported that *P. megacephala* was a single species within its exotic range.

## 2.2 | Molecular techniques

Genomic DNA was extracted from eight workers from each colony by using the Gentra Puregene Tissue kit (Qiagen) and stored at 4°C for subsequent genetic analyses. The DNA of these eight workers from each colony was used for sequencing and genotyping. The partial cytochrome oxidase I gene (COI) sequences, commonly used in DNA barcoding, were amplified using the primers LCO1490 (5'-GTCAACAAATCATAAA GATATTGG-3') and HCO2198 (5'-TAA ACTTCAGGGTGACCAAAAAATCA-3') to target a 708 bp fragment (Folmer et al., 1994). Polymerase chain reaction (PCR) was performed in a 25 µl reaction tube containing 2 µl of template DNA (25-50 ng), 12.5 µl of TaKaRa EmeraldAmp Max PCR Master Mix (TaKaRa), 0.2 μM forward and reverse primers, and ddH<sub>2</sub>O. The thermocycling conditions were as follows: initial denaturation at 94°C for 3 min, followed by 35 cycles, each consisting of 94°C for 30s, 55°C for 30s, and 72°C for 40s, and a final step at 72°C for 10 min. PCR products were cleaned using Zymo DNA Clean and Concentrator-5 Kit (Zymo Research), and were subjected to Sanger sequencing in both the forward and reverse directions.

Eight workers from each colony were genotyped, resulting in a total of 384 individuals, at seven dinucleotide-repeat microsatellite loci: Pmeg-06, Pmeg-07, Pmeg-09, Pmeg-10, Pmeg-11, Pmeg-12, and Pmeg-14 (Fournier et al., 2008). The PCR multiplex reactions were divided into two groups. The first group (Pmeg-06, Pmeg-09, Pmeg-12, and Pmeg-14) was subjected to the following thermocycling conditions: initial denaturation at 95°C for 15 min, followed by 35 cycles, each consisting of 94°C for 30s, 60°C for 90s, and 72°C for 60s, and a final step at 60°C for 30 min. The second group (Pmeg-07, Pmeg-10, and Pmeg-11) was subjected to the following thermocycling conditions: initial denaturation at 95°C for 15 min, followed by 35 cycles, each consisting of 94°C for 30 s, 56°C for 90 s, and 72°C for 60 s, and a final step at 60°C for 30 min. All PCR products were analyzed using the ABI 3730XL DNA Analyzer (Applied Biosystems) by Genomics BioSci and Tech (Taipei); GeneMarker (version 2.6.0; SoftGenetics LLC) was used to visualize and score alleles.

## 2.3 | Microsatellite genetic analyses in the Taiwanese population

GenAlEx (version 6.5; Peakall & Smouse, 2006) was used to evaluate the allele frequency, number of alleles  $N_{\rm A}$ , expected heterozygosity  $H_{\rm E}$ , and observed heterozygosity  $H_{\rm O}$  for every locus and region. The allelic richness Ar was calculated using FSTAT (Goudet, 2001). The genetic diversity of microsatellite loci was compared using analysis of variance conducted on SPSS (version 11.0; SPSS) followed by Tukey's honestly significant difference post hoc test for multiple comparisons ( $\alpha = .05$ );  $N_{\rm A}$  was compared using the Kruskal–Wallis test.

BOTTLENECK (Cornuet & Luikart, 1996; Piry et al., 1999) was employed to determine whether the population experienced a drastic reduction in genetic diversity by using distinct mutation models: a two-phase model (TPM) with 90% single-step mutations and another TPM with 10% multistep mutations. Significance was tested using a Wilcoxon signed-rank test as recommended by Piry et al. (1999) because the number of loci we used was <20. The occurrence of a bottleneck event was determined using the mode-shift test, which reveals the allele frequency distribution. Isolation by distance was determined by plotting the  $[F_{ST}/(1-F_{ST})]$  coefficients and the logarithm of the geographical distance. Isolation by distance was tested at among-colony and among-region levels. The significance of the correlation was tested using the Mantel test in GENEPOP (version 4.5; Raymond & Rousset, 1995).

Hierarchical analysis of molecular variance (AMOVA; Excoffier et al., 1992), implemented in GenAlEx (version 6.5; Peakall & Smouse, 2006), was used to determine variances at three levels: variances among the four geographical regions, variances among colonies, and variances between samples within colonies. The fixation indices for pairwise comparisons were determined using 999 permutations (Nei, 1973).

# 2.4 | Genetic differentiation among the three islands and within Taiwan

Genetic differentiation among Hawaii, Okinawa, and Taiwan, in addition to that among Taiwan's four regions, was estimated using  $F_{\rm ST}$ , and the corresponding significance was tested using a permutation test on GENEPOP (version 4.5; Raymond & Rousset, 1995). To investigate the population structures in Hawaii, Okinawa, and Taiwan, in addition to that in Taiwan's four regions, we used the Bayesian clustering method-based program STRUCTURE (version 2.3.4; Pritchard et al., 2000). All analyses were performed under an admixture model by using a Markov chain Monte Carlo (MCMC) run for 1 million generations with a 100,000 burn-in for cluster sizes ranging from 1 to 10 (i.e., K=1-10). Each K was tested 10 times. Structure Harvester (Earl & Vonholdt, 2012) was used to determine the best score for K values supported by the Delta K method (Evanno et al., 2005). The structure results were summarized and used to generate graphs through the CLUMPAK server

TABLE 1 Genetic diversity in seven microsatellite loci of Pheidole megacephala in Taiwan.

	TP (n = 48)	(1			TC (n = 64)				KH (n = 64)	1)			HT (n = 64)			
Z V		Ar	HE	н°	Z <sub>A</sub>	Ar	HE	Но	Z	Ar	HE	Η°	Z A	Ar	HE	H <sub>o</sub>
3.000		3.000	0.493	0.214	4.000	2.750	0.599	0.295	5.000	3.693	0.622	0.232	4.000	2.750	0.637	0.377
4.000		3.000	0.663	0.479	4.000	3.985	0.675	0.323	5.000	2.999	0.656	0.172	4.000	3.749	0.633	0.563
5.000		3.000	0.566	0.313	3.000	2.000	0.516	0.456	3.000	2.950	0.588	0.413	3.000	2.000	0.467	0.492
3.000		2.000	0.465	0.042	4.000	2.000	0.503	0.188	2.000	1.950	0.253	0.047	5.000	3.500	0.612	0.286
2.000		2.000	0.444	0.000	2.000	1.993	0.371	0.230	4.000	1.000	0.119	0.094	4.000	2.000	0.522	0.094
4.000		2.000	0.490	0.104	90009	3.699	0.668	0.397	9.000	2.000	0.487	0.234	9.000	4.250	0.684	0.323
2.000		2.000	0.153	0.125	5.000	1.999	0.443	0.311	2.000	2.000	0.492	0.469	4.000	2.943	0.519	0.339
3.286		2.430	0.468	0.182	4.000	2.632	0.540	0.314	3.857	2.370	0.460	0.237	4.286	3.027	0.582	0.353
	:	:	:								. :					

Abbreviations: Ar, allelic richness; H<sub>E</sub>, expected heterozygosity; H<sub>O</sub>, observed heterozygosity; n, total number of colonies; N<sub>A</sub>, number of alleles.

(http://clumpak.tau.ac.il; Jakobsson & Rosenberg, 2007). Principal component analysis (PCA) was used to visualize the population structure by plotting individual data in the R package adegenet (Jombart, 2008; R Core Team, 2020).

## 2.5 | mtDNA analyses

Sequences from both ends were manually edited and aligned using the ClustalW algorithm implemented in Bioedit (Hall, 2011) and assessed in MEGA (version 7.0; Kumar et al., 2016). Several COI sequences of P. megacephala from multiple studies (Fournier et al., 2012; Kartzinel & Pringle, 2015; Moreau, 2008; Smith & Fisher, 2009; Wills et al., 2014) were included in mtDNA analyses. Some of these mtDNA sequences do not share the exact genomic region amplified by the primers used in the present study; hence, only 310 bp fragments of these sequences that partially overlapped with our sequence were used in the subsequent analysis (Appendix S5). A phylogenetic tree based on truncated sequences was constructed using MrBayes (version 3.2; Ronquist et al., 2012) by selecting the generalized time-reversible model with gamma-distributed rate variation across sites and a proportion of invariable sites as the evolutionary model. Two parallel MCMC simulations were run for  $2 \times 10^6$ generations by using four chains (three heated and one cold), with each run sampling every 500 generations. A rapid bootstrap analysis and a search for the best-scoring maximum-likelihood tree were conducted using the extended majority rule-based bootstrapping criterion (Pattengale et al., 2010). All results were obtained using the general time-reversible nucleotide substitution model. Pheidole sexspinosa Mayr and P. xerophila (Fournier et al., 2012: Wheeler & Sauter, 1909) were included as outgroup species (Appendix S3).

## 3 | RESULTS

## 3.1 | Genetic diversity and bottleneck in the Taiwanese population

Two to six alleles were identified across seven polymorphic microsatellite loci (Table 1). A total of 12 private alleles across 6 of 7 microsatellite loci were identified. The average private allele frequencies were 0.049, 0.013, 0.008, and 0.035 in TP, TC, KH, and HT, respectively. The mean±standard error for  $N_{\rm A}$ , Ar,  $H_{\rm E}$ , and  $H_{\rm O}$  in Taiwan were  $3.857\pm0.234$ ,  $2.615\pm0.149$ ,  $0.512\pm0.028$ , and  $0.272\pm0.029$ , respectively. No significant differences in  $N_{\rm A}$ , Ar,  $H_{\rm E}$ , or  $H_{\rm O}$  were observed between the four regions (p=.509, .473, .351, and .148, respectively).

No significant heterozygosity excess was observed in all regions under the TPM model (Appendix S4 and Table S1). The allele frequency distribution revealed a high proportion of low-frequency alleles, resulting in a normal L-shaped curve in the mode-shift test (Figure 2). Because of the small sample size, a microsatellite data-fitting TPM model was used (Di Rienzo et al., 1994; Piry et al., 1999).

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The Wilcoxon signed-rank and mode-shift tests revealed that the populations had not experienced a recent bottleneck.

## 3.2 | Genetic differentiation and population structure within the Taiwanese population

Our AMOVA results (Table 2) revealed significant genetic structures at every hierarchical level; we observed 9% genetic variation among regions ( $F_{\rm RT}=0.085$ ), 43% among colonies ( $F_{\rm SR}=0.474$ ), and 48% within colonies ( $F_{\rm ST}=0.519$ ). The pairwise  $F_{\rm ST}$  values were moderate but significant across all combinations (Appendix S4 and Table S2), with the pairwise comparisons involving KH revealing high  $F_{\rm ST}$  values. We observed no significant positive correlations between geographical distance and genetic differentiation among the colonies and regions (Figure 3). It is worth noting that the isolation by distance is marginally significant in KH. The isolation by distance was not detected at the island scale most likely due to low sample size of the sampling regions (Figure 3b).

Our Bayesian clustering analysis conducted through STRUCTURE revealed that the optimal partitioning of all colony samples was K=2 (Figure 4). However, the population was not segregated into equivalent clusters on the basis of the four regions. The results indicated a weak population structure among the regions, suggesting a recent gene flow among these regions. As illustrated in Figure 4, most of the colonies sampled from TP and KH were present in clusters 1 (orange) and 2 (blue), respectively. The colonies sampled from TC and HT were partially observed in clusters 1 and 2 (Figure 4).

The PCA results revealed that the first two principal components accounted for 54.2% of the variation among Taiwan's four regions. The first principal component (44.8% of the total variance) mainly distinguished colonies in KH (southern Taiwan) from those in TP (northern Taiwan). The population structures of the colonies sampled from TC and HT, located in the central region of Taiwan, considerably overlapped with each other. Although the colonies sampled

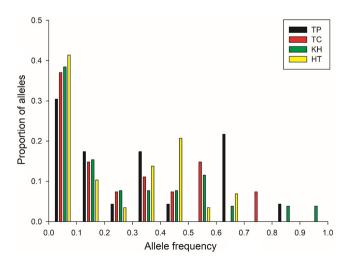


FIGURE 2 Allele frequency distribution at seven microsatellite loci in ants sampled from four regions.

from KH and TP partially overlapped with those sampled from TC and HT, the second principal component (9.4% of the total variance) indicated that the colonies sampled from KH possessed unique genetic structures distinct from those of the colonies sampled from the central region (TC and HT; Figure 5). Regarding the population structure, the PCA results were noted to be consistent with the STRUCTURE results.

## 3.3 | Genetic differentiation and population structures among the three islands

The pairwise  $F_{\rm ST}$  estimates demonstrated that the genetic differentiation between Okinawa and Hawaii was low ( $F_{\rm ST}=0.072$ ). However, the differentiation between Taiwan and Okinawa and that between Taiwan and Hawaii were moderate ( $F_{\rm ST}=0.176$  and 0.177, respectively; Table 3).

The PCA results revealed that the first two principal components accounted for 49.1% of the variation among islands (Figure 5). The first principal component (39.6% of the total variance) partially distinguished the populations in Hawaii and Okinawa from those in Taiwan. The Hawaiian and Okinawan populations were noted to almost overlap with each other, suggesting the genetic similarity between the two populations. The second principal component (9.5% of the total variance) could explain the genetic variation in the populations within the Okinawa and Hawaii regions. Our Bayesian clustering analysis revealed that the optimal partitioning of all colony samples was K = 2 (Figure 4). At K = 3, the populations from Hawaii and Okinawa exhibited similar genetic patterns. However, a specific genetic cluster was observed only in the Taiwanese population.

### 3.4 | Phylogenetic relationships

The phylogenetic tree (Figure 6)—constructed using 310bp of the mitochondrial control region—revealed that the Taiwanese population could be separated into three major clusters: Taiwan 1 (TW1), 2 (TW2), and 3 (TW3). TW1 and TW2 were separated by one mutational step (0.3%), whereas TW1 and TW3 were separated by seven mutational steps (2.3%). TW2 and TW3 were separated by six mutational steps (1.9%). Moreover, TW1 was most common haplotype (representing 77% of the individuals analyzed), followed by TW3 and TW2 (representing 20% and 3% of the individuals analyzed, respectively). TW1 was prevalent throughout Taiwan, whereas TW3 was noted in TC, KH, and HT but not in TP. By contrast, TW2 was noted only in KH (Figure 1).

TW3, belonging to the same phylogenetic group, resembles haplotypes recovered from specimens collected from the United States (Missouri and Florida), Australia, Cameroon, Mauritius, and Caribbean countries, whereas TW1 resembles haplotypes recovered from specimens collected from Hawaii and Okinawa (Figure 6). By contrast, TW2 represents a novel haplotype previously unreported in other regions.

Source	Df	MS	Percentage of variation (%)	F value	р
Among regions	3	38.701	9	$F_{RT} = 0.085$	.001
Among colonies	26	16.341	43	$F_{SR} = 0.474$	.001
Within colonies	450	1.058	48	$F_{ST} = 0.519$	.001
Total	479		100		

TABLE 2 AMOVA for Pheidole megacephala in Taiwan based on microsatellite data.

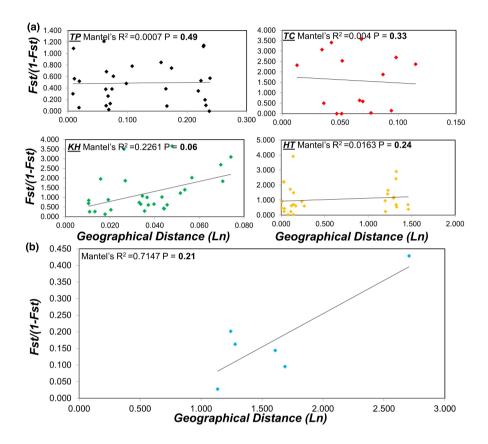


FIGURE 3 Correlations between genetic differentiation and geographical distances (isolation by distance) (a) among *Pheidole megacephala* colonies and (b) among Taiwanese regions.

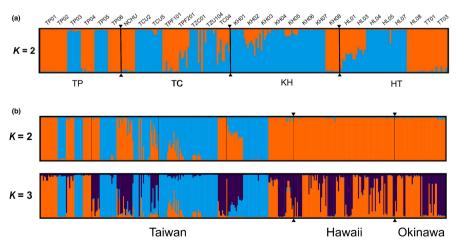


FIGURE 4 Population genetic structures based on Bayesian clustering analysis of *Pheidole megacephala* (a) among four Taiwanese regions (K = 2 for all sampled workers in all colonies) and (b) in Taiwan, Hawaii, and Okinawa (K = 2 and 3 for all sampled workers in all colonies).

### 4 | DISCUSSION

Our results demonstrate that *P. megacephala* in Taiwan has experienced a substantial reduction in genetic diversity. Comparisons of the genetic diversity of *P. megacephala* in Taiwan, Australia, and

South Africa (Fournier et al., 2009) indicated that Ar in Taiwan  $(2.615\pm0.149)$  is lower than that in South Africa  $(6.129\pm0.631)$  but similar to that in Australia  $(2.545\pm0.226)$ . Similar patterns of reduced genetic diversity have been commonly reported in the introduced ranges of most invasive ants (Fournier et al., 2005; Ross et al., 1993;

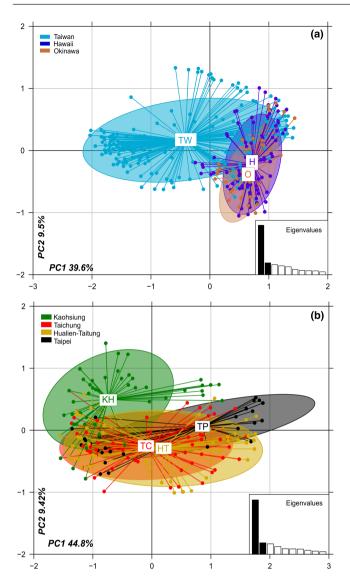


FIGURE 5 Principal component analysis clustering revealing the population structure across (a) four Taiwanese regions and (b) three Asia-Pacific islands.

TABLE 3 Pairwise genetic differentiation in the studied *Pheidole megacephala* populations in Taiwan, Hawaii, and Okinawa.

F <sub>ST</sub>	Taiwan	Hawaii	Okinawa
Taiwan	_	0.001	0.001
Hawaii	0.177	_	0.001
Okinawa	0.176	0.072	_

Vogel et al., 2010) and are best explained by population bottlenecks resulting from small founding populations (Arca et al., 2015; Colautti et al., 2017; Kinziger et al., 2011; Sakai et al., 2001).

Although we observed a lower Ar level in the *P. megacephala* populations in Taiwan than in those in South Africa, no evidence of heterozygosity excess was noted in the populations in Taiwan. The signature of heterozygosity excess is detectable in a bottlenecked population because the loss of alleles is typically faster than gene diversity during bottlenecks (Hedrick et al., 1986). The heterozygosity

excess method, however, captures only relatively recent bottlenecks occurring less than 0.2–4.0 adequate population size (Ne) generations ago (Luikart & Cornuet, 1998). Therefore, our finding likely reflects that the *P. megacephala* bottleneck during its introduction to Taiwan was not recent (Luikart et al., 1998). Reduced genetic diversity without a signature of heterozygosity excess was also observed in another introduced *P. megacephala* population of a similar age in Australia (Fournier et al., 2009). The heterozygosity excess method may detect a mutation–drift equilibrium in two introduced populations; however, its ability to detect a bottleneck can be affected by other factors (e.g., number of loci and population size), in addition to population recovery (Zepeda-Paulo et al., 2016).

Our analyses of Taiwan's P. megacephala population demonstrated that the levels of genetic differentiation among regions were significant but low (except in KH). We also observed no significant positive relationship between geographical and genetic distances (isolation by distance). Our finding corroborates those of most previous studies that have reported limited isolation by distance in the ant's introduced ranges. For example, a study observed no correlation between genetic and geographical distances of up to 3000km in Australia, which was engendered by a high level of gene flow between P. megacephala populations in human-dominating habitats (Fournier et al., 2009). These results are consistent with the predictions of an urban facilitation model of gene flow in other globally introduced Argentine ant populations in the United States, where urbanization unites ant populations through human-assisted longdistance dispersal (Tsutsui & Case, 2001). Nevertheless, a systematic sampling scheme, including a line transect across the island, may aid in understanding the dispersal dynamics of P. megacephala in Taiwan.

The bridgehead effect is a major pathway facilitating global ant invasion. Bertelsmeier et al. (2018) demonstrated that most P. megacephala populations intercepted at ports in the United States and New Zealand originated from the species' initial invasive populations. Similarly, recent port interception records revealed that all intercepted P. megacephala populations in Taiwan originated from a nonnative region (Lee, Weng, et al., 2020). As the spread of P. megacephala coincides with the first historical waves of globalization (Bertelsmeier et al., 2017), we speculated that the spread of P. megacephala to Taiwan might be linked to the historical trade openness during the Qing dynasty's colonization, during which four treaty ports-Keelung and Tamsui located in northern Taiwan, and Anping and Takow located in southern Taiwan-were opened to global trade. After 1860, the volume of good trades increased considerably (Yuju, 2017). To date, the Kaohsiung and Keelung ports are Taiwan's international commerce centers. This speculation is supported by our STRUCTURE results that the current distribution of P. megacephala may have resulted from two separate introduction events, one in TP (northern Taiwan) and the other in KH (southern Taiwan), both of which were most likely introduced from the ant's nonnative regions. Moreover, this speculation is supported by our phylogenetic analysis of P. megacephala collections from Taiwan compared with other populations from native and exotic ranges, suggesting two or three possible lineages.

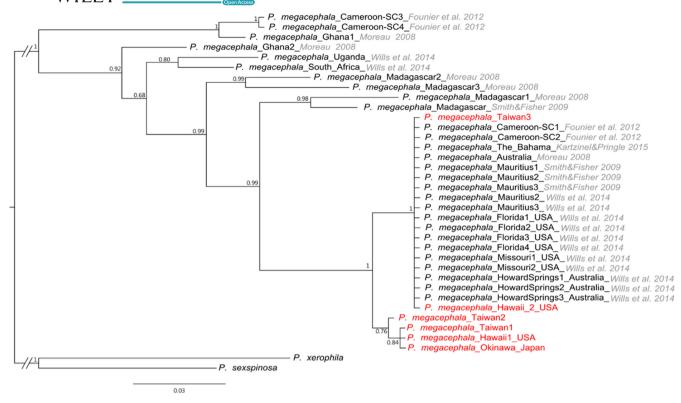


FIGURE 6 Bayesian inference tree based on 36 COI sequences (310 bp) of *Pheidole megacephala* from various localities with sister species—*Pheidole xerophila* and *Pheidole sexspinosa*—as outgroups. Red text denotes haplotypes identified in this study.

Haplotype TW1, which shares a common origin with introduced populations in Okinawa and Hawaii, was responsible for the regionwide distribution in Taiwan. These P. megacephala populations may have arisen multiple times from unknown origins, considering that research on the phylogenetics of the test species is limited. Additionally, the PCA and STRUCTURE results indicate an apparent nuclear genetic similarity among the three P. megacephala populations. All of these results suggest a high level of connectivity among the three islands and that the interisland spread of P. megacephala might have been expected. This suggestion is supported by the extensive human immigration events occurring in Japan, Taiwan (i.e., Okinawa and Taiwan; Taiwan was under Japanese rule during the period under consideration), and Hawaii during the late 19th and early 20th centuries (Boyd, 1971; Matsumoto, 1982). Considering the ant's hitchhiker-like nature, we speculated that P. megacephala might have been frequently transported among these regions through these immigration events. A similar genetic pattern was also reported for another invasive ant, Anoplolepis gracilipes (yellow crazy ant), in southern Japan, Taiwan, and Hawaii (Lee, Lin, et al., 2020), reinforcing the role of human-assisted jump dispersal in shaping the genetic structure of invasive ants.

We observed a previously unreported haplotype, namely TW2. TW2 was limited to KH and was identified in a park 3 km from an international seaport (Port of Kaohsiung). In 2018, 100 colonies of imported red fire ants were discovered in the container yard at the Port of Kaohsiung (received cargo from China (60%) and the United States (40%); Wylie et al., 2020). The port of Kaohsiung is one of the

busiest and largest seaports in the world, ranking 15th for international container shipping (Lauriat, 2019; Niimi, 2004). The low occurrence of TW2 (12.5%) and its limited distribution in KH suggest that the introduction of individuals harboring this haplotype occurred recently. Additional comprehensive studies on *P. megacephala* populations from other Asia-Pacific regions are necessary to clarify their invasion history.

### **AUTHOR CONTRIBUTIONS**

Kuan-Ling Liu: Conceptualization (supporting); formal analysis (lead); investigation (lead); visualization (lead); writing – original draft (lead); writing – review and editing (equal). Shu-Ping Tseng: Methodology (equal); writing – review and editing (equal). Haruki Tatsuta: Resources (supporting); writing – review and editing (equal). Kazuki Tsuji: Resources (supporting); writing – review and editing (equal). Jia-Wei Tay: Resources (supporting); writing – review and editing (supporting). G. Veera Singham: Validation (equal); writing – review and editing (equal). Chin-Cheng Scotty Yang: Conceptualization (supporting); methodology (supporting); validation (equal); writing – review and editing (equal). Kok Boon Neoh: Conceptualization (supporting); funding acquisition (lead); supervision (lead); writing – original draft (lead); writing – review and editing (equal).

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### CONFLICT OF INTEREST

The authors declare no competing financial interests.

#### DATA AVAILABILITY STATEMENT

All raw sequence files have been deposited in the National Center for Biotechnology Information (NCBI) under GenBank accession numbers ON524410-ON524414 and ON528936.

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#### REFERENCES

- Arca, M., Mougel, F., Guillemaud, T., Dupas, S., Rome, Q., Perrard, A., Muller, F., Fossoud, A., Capdevielle-Dulac, C., Torres-Leguizamon, M., Chen, X. X., Tan, J. L., Jung, C., Villemant, C., Arnold, G., & Silvain, J. F. (2015). Reconstructing the invasion and the demographic history of the yellow-legged hornet, Vespa velutina, in Europe. Biological Invasions, 17(8), 2357–2371. https://doi.org/10.1007/s10530-015-0880-9
- Ascunce, M. S., Yang, C. C., Oakey, J., Calcaterra, L., Wu, W. J., Shih, C. J., Goudet, J., Ross, K. G., & Shoemaker, D. W. (2011). Global invasion history of the fire ant *Solenopsis invicta*. *Science*, 331(6020), 1066– 1068. https://doi.org/10.1126/science.1198734
- Bertelsmeier, C., Ollier, S., Liebhold, A., & Keller, L. (2017). Recent human history governs global ant invasion dynamics. *Nature Ecology & Evolution*, 1(7), 0184. https://doi.org/10.1038/s41559-017-0184
- Bertelsmeier, C., Ollier, S., Liebhold, A. M., Brockerhoff, E. G., Ward, D., & Keller, L. (2018). Recurrent bridgehead effects accelerate global alien ant spread. *Proceedings of the National Academy of Sciences*, 115(21), 5486–5491. https://doi.org/10.1073/pnas.1801990115
- Blumenfeld, A. J., Eyer, P. A., Husseneder, C., Mo, J., Johnson, L. N. L., Wang, C., Kenneth Grace, J., Chouvenc, T., Wang, S., & Vargo, E. L. (2021). Bridgehead effect and multiple introductions shape the global invasion history of a termite. *Communications Biology*, 4(1), 196. https://doi.org/10.1038/s42003-021-01725-x
- Bolton, B. (1994). *Identification guide to the ant genera of the world.*Harvard University Press.
- Boyd, M. (1971). Oriental immigration: The experience of the Chinese, Japanese, and Filipino populations in the United States. *International Migration Review*, 5(1), 48–61.
- Burwell, C. J., Nakamura, A., McDougall, A., & John Neldner, V. (2012). Invasive African big-headed ants, *Pheidole megacephala*, on coral cays of the southern great barrier reef: Distribution and impacts on other ants. *Journal of Insect Conservation*, 16(5), 777–789. https://doi.org/10.1007/s10841-012-9463-6
- Callan, S. K., & Majer, J. D. (2009). Impacts of an incursion of African bigheaded ants, *Pheidole megacephala* (Fabricius), in urban bushland in Perth, Western Australia. *Pacific Conservation Biology*, 15(2), 102–115. https://doi.org/10.1071/PC090102
- Chang, V. (1985). Colony revival, and notes on rearing and life history of the big-headed ant. *Proceedings of the Hawaiian Entomological Society*, 25, 53-58.
- Colautti, R. I., Alexander, J. M., Dlugosch, K. M., Keller, S. R., & Sultan, S. E. (2017). Invasions and extinctions through the looking glass of evolutionary ecology. *Philosophical Transactions of the Royal Society, B: Biological Sciences*, 372(1712), 20160031. https://doi.org/10.1098/rstb.2016.0031

- Cornuet, J. M., & Luikart, G. (1996). Description and power analysis of two tests for detecting recent population bottlenecks from allele frequency data. *Genetics*, 144(4), 2001–2014.
- Dejean, A., Kenne, M., & Moreau, C. S. (2007). Predatory abilities favour the success of the invasive ant *Pheidole megacephala* in an introduced area. *Journal of Applied Entomology*, 131(9–10), 625–629. https://doi.org/10.1111/j.1439-0418.2007.01223.x
- Dejean, A., Moreau, C. S., Kenne, M., & Leponce, M. (2008). The raiding success of *Pheidole megacephala* on other ants in both its native and introduced ranges. *Comptes Rendus Biologies*, 331(8), 631–635.
- Di Rienzo, A., Peterson, A. C., Garza, J. C., Valdes, A. M., Slatkin, M., & Freimer, N. B. (1994). Mutational processes of simple-sequence repeat loci in human populations. *Proceedings of the National Academy of Sciences*, 91(8), 3166–3170. https://doi.org/10.1073/pnas.91.8.3166
- Drescher, J., Blüthgen, N., Schmitt, T., Bühler, J., & Feldhaar, H. (2010). Societies drifting apart? Behavioural, genetic and chemical differentiation between supercolonies in the yellow crazy ant Anoplolepis gracilipes. PLoS One, 5(10), e13581. https://doi.org/10.1371/journal.pone.0013581
- Earl, D. A., & Vonholdt, B. M. (2012). STRUCTURE HARVESTER: A website and program for visualizing STRUCTURE output and implementing the Evanno method. *Conservation Genetics Resources*, 4(2), 359–361. https://doi.org/10.1007/s12686-011-9548-7
- Evanno, G., Regnaut, S., & Goudet, J. (2005). Detecting the number of clusters of individuals using the software structure: A simulation study. *Molecular Ecology*, 14(8), 2611–2620. https://doi.org/10.1111/j.1365-294X.2005.02553.x
- Excoffier, L., Smouse, P. E., & Quattro, J. M. (1992). Analysis of molecular variance inferred from metric distances among DNA haplotypes: Application to human mitochondrial DNA restriction data. *Genetics*, 131(2), 479–491. https://doi.org/10.1093/genetics/131.2.479
- Eyer, P.-A., Matsuura, K., Vargo, E. L., Kobayashi, K., Yashiro, T., Suehiro, W., Himuro, C., Yokoi, T., Guénard, B., Dunn, R. R., & Tsuji, K. (2018). Inbreeding tolerance as a pre-adapted trait for invasion success in the invasive ant *Brachyponera chinensis*. *Molecular Ecology*, 27(23), 4711–4724. https://doi.org/10.1111/mec.14910
- Ficetola, G. F., Bonin, A., & Miaud, C. (2008). Population genetics reveals origin and number of founders in a biological invasion. *Molecular Ecology*, 17(3), 773–782.
- Folmer, O., Black, M., Hoeh, W., Lutz, R., & Vrijenhoek, R. (1994). DNA primers for amplification of mitochondrial cytochrome coxidase subunit I from diverse metazoan invertebrates. *Molecular Marine Biology and Biotechnology*, 3(5), 294–299.
- Fournier, D., De Biseau, J. C., & Aron, S. (2009). Genetics, behaviour and chemical recognition of the invading ant *Pheidole megacephala*. *Molecular Ecology*, 18(2), 186–199.
- Fournier, D., Dubois, D., & Aron, S. (2008). Isolation and characterization of microsatellite loci from the invasive ant *Pheidole megacephala*. *Molecular Ecology Resources*, 8(4), 919–922.
- Fournier, D., Foucaud, J., Loiseau, A., Cros-Arteil, S., Jourdan, H., Orivel, J., Le Breton, J., Chazeau, J., Dejean, A., Keller, L., & Estoup, A. (2005). Characterization and PCR multiplexing of polymorphic microsatellite loci for the invasive ant Wasmannia auropunctata. Molecular Ecology Notes, 5(2), 239–242.
- Fournier, D., Tindo, M., Kenne, M., Mbenoun Masse, P. S., Van Bossche, V., De Coninck, E., & Aron, S. (2012). Genetic structure, nestmate recognition and behaviour of two cryptic species of the invasive big-headed ant *Pheidole megacephala*. PLoS One, 7(2), e31480.
- Garnas, J. R., Auger-Rozenberg, M.-A., Roques, A., Bertelsmeier, C., Wingfield, M. J., Saccaggi, D. L., Roy, H. E., & Slippers, B. (2016). Complex patterns of global spread in invasive insects: Ecoevolutionary and management consequences. *Biological Invasions*, 18(4), 935–952.

- Goudet, J. (2001). FSTAT, a program to estimate and test gene diversities and fixation indices, version 2.9.3. http://www2unilch/popgen/softwares/fstathtm
- Hall, T. (2011). BioEdit: An important software for molecular biology. GERF Bulletin of Bioscience, 2, 60–61.
- Hedrick, P. W., Brussard, P. F., Allendorf, F. W., Beardmore, J. A., & Orzack, S. (1986). Protein variation, fitness, and captive propagation. *Zoo Biology*, 5(2), 91–99. https://doi.org/10.1002/zoo.14300 50204
- Hee, J. J., Holway, D. A., Suarez, A. V., & Case, T. J. (2000). Role of propagule size in the success of incipient colonies of the invasive argentine ant. *Conservation Biology*, 14(2), 559–563. https://doi.org/10.1046/j.1523-1739.2000.99040.x
- Hoffmann, B. D. (2014). Quantification of supercolonial traits in the yellow crazy ant, *Anoplolepis gracilipes*. *Journal of Insect Science*, 14(1), 25. https://doi.org/10.1093/jis/14.1.25
- Hoffmann, B. D., Andersen, A. N., & Hill, G. J. (1999). Impact of an introduced ant on native rain forest invertebrates: *Pheidole megacephala* in monsoonal Australia. *Oecologia*, 120(4), 595–604. https://doi.org/10.1007/PL00008824
- Hulme, P. E. (2009). Trade, transport and trouble: Managing invasive species pathways in an era of globalization. *Journal of Applied Ecology*, 46(1), 10–18.
- Jakobsson, M., & Rosenberg, N. A. (2007). CLUMPP: A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. *Bioinformatics*, 23(14), 1801–1806. https://doi.org/10.1093/bioinformatics/btm233
- Jombart, T. (2008). Adegenet: A R package for the multivariate analysis of genetic markers. *Bioinformatics*, 24(11), 1403–1405. https://doi. org/10.1093/bioinformatics/btn129
- Kartzinel, T. R., & Pringle, R. M. (2015). Molecular detection of invertebrate prey in vertebrate diets: Trophic ecology of Caribbean Island lizards. Molecular Ecology Resources, 15(4), 903–914. https://doi. org/10.1111/1755-0998.12366
- Kinziger, A. P., Nakamoto, R. J., Anderson, E. C., & Harvey, B. C. (2011). Small founding number and low genetic diversity in an introduced species exhibiting limited invasion success (speckled dace, Rhinichthys osculus). Ecology and Evolution, 1(1), 73–84. https://doi. org/10.1002/ece3.8
- Kumar, S., Stecher, G., & Tamura, K. (2016). MEGA7: Molecular evolutionary genetics analysis version 7.0 for bigger datasets. Molecular Biology and Evolution, 33(7), 1870–1874. https://doi.org/10.1093/molbev/msw054
- Lauriat, G. (2019). Top 100 global container ports: AJOT's top 100 containerports A to Z. American Journal of Transportation. http://www.ajot.com/premium/ajot-ajots-top-100-containerports-a-to-z/P0. Accessed March 30, 2021.
- Lee, C. C., Lin, C. Y., Tseng, S. P., Matsuura, K., & Yang, C. C. S. (2020). Ongoing coevolution of Wolbachia and a widespread invasive ant, *Anoplolepis gracilipes. Microorganisms*, 8(10), 1569.
- Lee, C. C., Weng, Y. M., Lai, L. C., Suarez, A. V., Wu, W. J., Lin, C. C., & Yang, C. C. S. (2020). Analysis of recent interception records reveals frequent transport of arboreal ants and potential predictors for ant invasion in Taiwan. *Insects*, 11(6), 356. https://doi.org/10.3390/insects11060356
- Lin, C. C. (1998). Systematic and zoogeographic studies on the ant subfamily Myrmicinae in Taiwan (Hymenoptera: Formicidae). [Doctoral dissertation, PhD Dissertation, National Taiwan University Press, Taiwan].
- Liu, K. L. (2020). Population structure and colony organization of the invasive big-headed ant (Pheidole megacephala) in Taiwan. [Master's Thesis, National Chung-Hsing University].
- Lombaert, E., Guillemaud, T., Cornuet, J.-M., Malausa, T., Facon, B., & Estoup, A. (2010). Bridgehead effect in the worldwide invasion of the biocontrol harlequin ladybird. *PLoS One*, 5(3), e9743.
- Luikart, G., Allendorf, F., Cornuet, J. M., & Sherwin, W. (1998). Distortion of allele frequency distributions provides a test for recent

- population bottlenecks. *Journal of Heredity*, 89(3), 238–247. https://doi.org/10.1093/jhered/89.3.238
- Luikart, G., & Cornuet, J. M. (1998). Empirical evaluation of a test for identifying recently bottlenecked populations from allele frequency data. Conservation Biology, 12, 228–237.
- Matsumoto, Y. S. (1982). Okinawa migrants to Hawaii. *Hawaiian Journal of History*, 16, 125–133.
- Meyerson, L. A., & Mooney, H. A. (2007). Invasive alien species in an era of globalization. *Frontiers in Ecology and the Environment*, 5(4), 199–208. https://doi.org/10.1890/1540-9295(2007)5[199:IASIA E]2.0.CO:2
- Moreau, C. S. (2008). Unraveling the evolutionary history of the hyperdiverse ant genus Pheidole (Hymenoptera: Formicidae). Molecular Phylogenetics and Evolution, 48(1), 224–239. https://doi.org/10.1016/j.ympev.2008.02.020
- Nei, M. (1973). Analysis of gene diversity in subdivided populations. Proceedings of the National Academy of Sciences, 70(12), 3321–3323.
- Niimi, A. J. (2004). Role of container vessels in the introduction of exotic species. Marine Pollution Bulletin, 49(9-10), 778-782. https://doi. org/10.1016/j.marpolbul.2004.06.006
- Pattengale, N. D., Alipour, M., Bininda-Emonds, O. R., Moret, B. M., & Stamatakis, A. (2010). How many bootstrap replicates are necessary? *Journal of Computational Biology*, 17(3), 337–354. https://doi.org/10.1089/cmb.2009.0179
- Peakall, R., & Smouse, P. E. (2006). Genalex 6: Genetic analysis in excel. Population genetic software for teaching and research. *Molecular Ecology Notes*, 6(1), 288–295. https://doi.org/10.1111/j.1471-8286.2005.01155.x
- Piry, S., Luikart, G., & Cornuet, J. M. (1999). BOTTLENECK: A computer program for detecting recent recent effective population size reductions from allele data frequencies. *Journal of Heredity*, 90(4), 502–503. https://doi.org/10.1093/jhered/90.4.502
- Plentovich, S., Hebshi, A., & Conant, S. (2009). Detrimental effects of two widespread invasive ant species on weight and survival of colonial nesting seabirds in the Hawaiian islands. *Biological Invasions*, 11(2), 289–298.
- Pritchard, J. K., Stephens, M., & Donnelly, P. (2000). Inference of population structure using multilocus genotype data. *Genetics*, 155(2), 945–959
- R Core Team. (2020). R: A language and environment for statistical computing. R Foundation for Statistical Computing.
- Raymond, M., & Rousset, F. (1995). GENEPOP (version 1.2): Population genetics software for exact tests and ecumenicism. *Journal of Heredity*, 86, 248–249. https://doi.org/10.1093/oxfordjournals.jhered.a111573
- Ronquist, F., Teslenko, M., van der Mark, P., Ayres, D. L., Darling, A., Höhna, S., Larget, B., Liu, L., Suchard, M. A., & Huelsenbeck, J. P. (2012). MrBayes 3.2: Efficient Bayesian phylogenetic inference and model choice across a large model space. *Systematic Biology*, *61*(3), 539–542. https://doi.org/10.1093/sysbio/sys029
- Ross, K., Vargo, E., Keller, L., & Trager, J. (1993). Effect of a founder event on variation in the genetic sex-determining system of the fire ant *Solenopsis invicta*. *Genetics*, 135(3), 843–854.
- Sakai, A. K., Allendorf, F. W., Holt, J. S., Lodge, D. M., Molofsky, J., With, K. A., Baughman, S., Cabin, R. J., Cohen, J. E., Ellstrand, N. C., McCauley, D. E., O'Neil, P., Parker, I. M., Thompson, J. N., & Weller, S. G. (2001). The population biology of invasive species. *Annual Review of Ecology and Systematics*, 32(1), 305–332. https://doi.org/10.1146/annurev.ecolsys.32.081501.114037
- Sarnat, E. M., Fischer, G., Guénard, B., & Economo, E. P. (2015). Introduced Pheidole of the world: Taxonomy, biology and distribution. *ZooKeys*, 543, 1–109.
- Sherpa, S., Rioux, D., Goindin, D., Fouque, F., François, O., & Després, L. (2017). At the origin of a worldwide invasion: Unraveling the genetic makeup of the Caribbean bridgehead populations of the

- dengue vector Aedes aegypti. Genome Biology and Evolution, 10(1), 56-71. https://doi.org/10.1093/gbe/evx267
- Smith, M. A., & Fisher, B. L. (2009). Invasions, DNA barcodes, and rapid biodiversity assessment using ants of Mauritius. *Frontiers in Zoology*, 6(1), 1–12. https://doi.org/10.1186/1742-9994-6-31
- Strohecker, L. F. (2012). Ants and Hawaiian seabirds A deadly combination. Maui Invasive Species Committee. https://mauiinvasive.org/2012/11/09/ants-and-hawaiian-seairds-a-totally-unnatural-combination/
- Suarez, A. V., & Tsutsui, N. D. (2008). The evolutionary consequences of biological invasions. *Molecular Ecology*, 17(1), 351–360.
- Tay, J. W., Neoh, K. B., & Lee, C. Y. (2014). The roles of the queen, brood, and worker castes in the colony growth dynamics of the pharaoh ant, *Monomorium pharaonis* (Hymenoptera: Formicidae). *Myrmecological News*, 20, 87–94.
- Tsai, C. Y. (2019). Diversity, community structure and morphological patterns of ground-dwelling ant in urban-rural interface. [Master's thesis, National Chung-Hsing University].
- Tsutsui, N. D., & Case, T. J. (2001). Population genetics and colony structure of the argentine ant (linepithema humile) in its native and introduced ranges. *Evolution*, *55*(5), 976–985. https://doi.org/10.1111/j.0014-3820.2001.tb00614.x
- Tsutsui, N. D., Suarez, A. V., Holway, D. A., & Case, T. J. (2000). Reduced genetic variation and the success of an invasive species. *Proceedings of the National Academy of Sciences*, *97*(11), 5948–5953.
- Vanderwoude, C., Lobry De Bruyn, L. A., & House, A. P. (2000). Response of an open-forest ant community to invasion by the introduced ant, *Pheidole megacephala*. *Austral Ecology*, 25(3), 253–259. https://doi.org/10.1046/j.1442-9993.2000.01021.x
- Vogel, V., Pedersen, J. S., Giraud, T., Krieger, M. J., & Keller, L. (2010). The worldwide expansion of the argentine ant. Diversity and Distributions, 16(1), 170-186.
- Wetterer, J. K. (2007). Biology and impacts of Pacific Island invasive species. 3. The African big-headed ant, *Pheidole megacephala* (Hymenoptera: Formicidae). *Pacific Science*, 61(4), 437–457.
- Wetterer, J. K. (2012). Worldwide spread of the African big-headed ant, Pheidole megacephala (Hymenoptera: Formicidae). Myrmecological News, 17, 51–62.

- Wheeler, W. M., & Sauter, H. (1909). Ants of Formosa and The Philippines.

  Bulletin of the American Museum of Natural History, 26, 333–345.
- Wills, B. D., Moreau, C. S., Wray, B. D., Hoffmann, B. D., & Suarez, A. V. (2014). Body size variation and caste ratios in geographically distinct populations of the invasive big-headed ant, *Pheidole megacephala* (Hymenoptera: Formicidae). *Biological Journal of the Linnean Society*, 113(2), 423–438. https://doi.org/10.1111/bii.12386
- Wylie, R., Yang, C. C. S., & Tsuji, K. (2020). Invader at the gate: The status of red imported fire ant in Australia and Asia. *Ecological Research*, 35(1), 6–16.
- Yuju, L. (2017). Continuity and breakdown: Taiwan's customs service during the Japanese occupation, 1895–1945. *International Journal* of Maritime History, 29(4), 855–874. https://doi.org/10.1177/08438 71417726967
- Zepeda-Paulo, F., Dion, E., Lavandero, B., Mahéo, F., Outreman, Y., Simon, J. C., & Figueroa, C. C. (2016). Signatures of genetic bottle-neck and differentiation after the introduction of an exotic parasit-oid for classical biological control. *Biological Invasions*, 18(2), 565–581. https://doi.org/10.1007/s10530-015-1029-6

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